Use of iron oxide nanomaterials as adsorbents for wastewater treatment

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Abstract – The dissemination of chemical components into the environment might provoke harm to human health, animal species and the ecosystem. In order to reduce the negative effects of those pollutants, physico-chemical treatment procedures, namely the adsorption technique has been implemented. This study seeks to synthesize manufactured nanomaterials throughout a co-precipitation process such as nanoparticles (Ferrihydrite (Fh)) and nanocomposite magnetic iron (magnetite / activated carbon) capable of adsorbing certain inorganic species (heavy metals) such as hexavalent chromium Cr (VI). The prepared supports were characterized by IRTF analysis. The influence of the various experimental parameters (the initial concentration and the contact time) have been studied. Various isothermal models have been applied, namely Langmuir, Freundlich, Temkin and BET. The study of those absorption isotherms has shown that Langmuir’s model describes better Cr (VI) adsorption. It has been also noticed that the adsorption process on the Fh follows a pseudo 2nd order kinetics for the hexavalent chromium. The experimental results have proved that equilibrium is reached after 15 minutes with a better yield of 99.511% and a maximum retention capacity of 40.9157 mg / g.

Keywords – Ferrihydrite (Fh); Magnetite; Manufactured Nanomaterials; Nanocomposite; Co-Precipitation; Isotherm

I. INTRODUCTION

The release of chemical compounds into the environment can cause damage to human health, animal species and the ecosystem. In order to reduce the negative effects of these pollutants, physico-chemical treatment procedures, including the adsorption technique, have been implemented [1]. This study aims at synthesizing engineered nanomaterials through a co-precipitation process namely; Ferrihydrite (Fh) nanoparticles and magnetic iron nanocomposites (magnetite/activated carbon) capable of adsorbing heavy metals such as hexavalent chromium Cr (VI).

II. MATERIALS AND METHOD

The approach that we followed was to test the adsorption capacity and effectiveness of particles (Ferrihydrite) for the elimination and depollution of industrial effluent containing metallic inorganic pollutants (Cr (VI)). The current study is organized around two main points:
Chemical co-precipitation is used to synthesize adsorbents, which are manufactured nanoparticles such as ferrihydrite and composite materials based on iron oxide and industrial activated carbon. Then, they were characterized by IRTF analysis.

Evaluation of adsorption experiments (UV-Visible spectrophotometric dose) of the mineral contaminant hexavalent chromium on synthetic ferrihydrite powder and composite material.

III. RESULTS

ANALYSIS BY IRTF SPECTROSCOPY

![IR spectrum of the synthesized Ferrihydrite (Fh) before (a) and after adsorption (b)]

Table 1. Main detected groups of Fh before adsorption

<table>
<thead>
<tr>
<th>Wave number (cm$^{-1}$)</th>
<th>Bonds</th>
<th>Nature of the vibration</th>
</tr>
</thead>
<tbody>
<tr>
<td>3416.28</td>
<td>O-H or N-H</td>
<td>Elongation (stretching) vibration of the hydroxyl groups (alcohol) of the O-H bonds or with the N-H amine groups [2], [3], [4].</td>
</tr>
<tr>
<td>1626.66</td>
<td>H$_2$O</td>
<td>Water molecules adsorbed on nanoparticles [5], [6], [7].</td>
</tr>
<tr>
<td>1531.20 and 1481.06</td>
<td>Fe$_3$O</td>
<td>Vibration band corresponding to the deformation (bending) of Fe-O links [7], [6].</td>
</tr>
<tr>
<td>1400.07</td>
<td>Fe$_3$-OH</td>
<td>Vibration band corresponding to the deformation of Fe-OH bonds [6], [7], [8].</td>
</tr>
</tbody>
</table>

Fig. 1 Synthesis of iron oxide nanoparticles a) Ferrihydrite (Fh); b) Nanocomposite (Magnetite/Active Carbon (CA))

Fig. 2 IR spectrum of the synthesized Ferrihydrite (Fh) before (a) and after adsorption (b)
Table 2. Main detected groups of Fh after adsorption

<table>
<thead>
<tr>
<th>Wave number (cm$^{-1}$)</th>
<th>Bonds</th>
<th>Nature of the vibration</th>
</tr>
</thead>
<tbody>
<tr>
<td>3582.13 (Very wide)</td>
<td>O-H or N-H</td>
<td>The absorption band is probably associated with the elongation (stretching) vibration of the hydroxyl groups (alcohol) of the O-H bonds or with the N-H amine groups [9], [10].</td>
</tr>
<tr>
<td>1716.34</td>
<td>H$_2$O</td>
<td>Absorption band corresponding to water molecules adsorbed on nanoparticles [6].</td>
</tr>
<tr>
<td>1544.70, 1491.67, 1474.31 and 1435.74</td>
<td>Fe-O Or Fe-OH</td>
<td>Vibration band corresponding to the deformation (bending) of Fe-O or Fe-OH links [6].</td>
</tr>
</tbody>
</table>

Fig. 3 IR spectrum of the synthesized (Fe$_3$O$_4$/ CA) composite

2. EFFECT OF CONTACT TIME

Fig. 4 The Effect of contact time of Cr (VI) by Fh
Conditions : $C_0 = 50$ mg/L, $T= 25^\circ$C, $d= 0.125$mm, $v= 400$ tr/min, $r$ (S/L) = 10 g/L

3. EFFECT OF THE INITIAL CONCENTRATION

Fig. 5 The effect of initial concentration on Cr (VI) removal by Fh.
Conditions : $T= 25^\circ$C, $d= 0.125$mm, $v= 400$ tr/min, $r$ (S/L) = 10 g/L
4. MODELIZATION

<table>
<thead>
<tr>
<th>Constants</th>
<th>C₀=50 mg/L</th>
<th>C₀=100 mg/L</th>
<th>C₀=200 mg/L</th>
<th>C₀=400 mg/L</th>
</tr>
</thead>
<tbody>
<tr>
<td>K₂ (g. mg⁻¹ min⁻¹)</td>
<td>2.122</td>
<td>6.682</td>
<td>19.801</td>
<td>0.5000</td>
</tr>
<tr>
<td>qₑ (mg/g)</td>
<td>4.924</td>
<td>9.794</td>
<td>19.801</td>
<td>39.840</td>
</tr>
<tr>
<td>R²</td>
<td>0.999</td>
<td>1</td>
<td>1</td>
<td>1</td>
</tr>
</tbody>
</table>

5. ADSORPTION’S ISOTHERMS

<table>
<thead>
<tr>
<th>Types of isotherms</th>
<th>Isotherm’s constants</th>
<th>R²</th>
</tr>
</thead>
<tbody>
<tr>
<td>Langmuir</td>
<td>qₘ=40.9157 (mg/g) Kₐ=0.64635 (L/mg)</td>
<td>0.9495</td>
</tr>
<tr>
<td>Freundlich</td>
<td>Kᵣ=16.3592 (mg.g⁻¹)(L.mg⁻¹)¹/n</td>
<td>n= 1.2122</td>
</tr>
<tr>
<td>Temkin</td>
<td>bᵣ=0.2031654 KJ/mol</td>
<td>Ln Aᵣ= 1.7160</td>
</tr>
<tr>
<td>BET</td>
<td>qₘ=13.1788 (mg/g) K= 850.718</td>
<td>0.46915</td>
</tr>
</tbody>
</table>

IV. DISCUSSION

IV.1. ANALYSIS BY IRTF SPECTROSCOPY

According to Figure 2 and Tables 1 and 2, we observed that when the spectra of the two samples, manufactured ferrihydrite (Fh) (before adsorption) and ferrihydrite loaded with hexavalent chromium (after adsorption), are compared, a shift in practically all peaks is seen, with an increase in the strength of the existing peaks. This shows that functional groups (carboxyl, alcohol, phenol, and amine) are involved in the adsorption of Cr (VI) on Fh.

The IR spectrum of the synthesized composite (Fe₃O₄ / CA) is shown in this figure (Fig. 3) an adsorbed water produces a weak peak at 3423.03 cm⁻¹. While the peaks at 3389.28 cm⁻¹ and 1116.58 cm⁻¹ are caused by stretching and bending vibrations of the O-H bonds, which support the existence of hydroxyl functional groups, respectively (which may be an alcohol, phenol, or carboxyl) [11], [12], [13] and a modest peak at 1116.58 cm⁻¹ indicates the stretching vibration of C=OOH bonds, but the peak at 1330.64 cm⁻¹ in the region (1400–1000 cm⁻¹) indicates the elongation vibration of C-OH bonds, confirming the existence of -COOH groups[14]. Indeed, the elongation vibrations of the C = C and C - C bonds are shown by two peaks at 1625.12 cm⁻¹ and 839.84 cm⁻¹, respectively. At 566.969 cm⁻¹, a wide band of extremely high intensity is tuned to the stretching vibration of Fe - O bonds in the crystal lattice [13].

IV.2. EFFECT OF CONTACT TIME

According to figure 4 (Fig. 4), which represents the variation of the adsorption capacity q as a function of the contact time, we noticed that equilibrium is reached after 15 minutes with an efficiency of 99.511%, indicating the presence of two phases: one rapid and one constant. Also, the speed decreases until 60 minutes have passed, indicating that the sites are saturated.
IV.3. EFFECT OF THE INITIAL CONCENTRATION

In accordance with the figure (Fig. 5), we noticed a proportionality between the retention capacity and the initial concentration of Cr (VI). Indeed, an increase in the amount of pollutant adsorbed on unbound active sites. The maximal retention capacity rises from 4.9163 mg/g for a 50 mg/L starting dosage to 39.8075 mg/g for a 400 mg/L initial concentration.

IV.4. MODELIZATION

According to the recapitulative table (Table 3), we observed that the \( R^2 \) numbers for the pseudo-second-order model are very satisfactory. It can be inferred that the pseudo-2nd-order model better describes the retention kinetics of Cr (VI) by Fh than the other models.

IV.5. ADSORPTION’S ISOTHERMS

When compared to other models (Table 4), the Cr (VI) retention mechanism is best described by a single-layer monolayer isotherm (the Langmuir isotherm) with the greatest correlation factor and the closest to unity (indicating adequate linearization (Langmuir isotherm \( q_m = 40.9157 \) mg/g)). Indeed, there is no discernible difference between the Freundlich and Langmuir \( R^2 \) numbers. We also observe that \( n \) between 0 and 10 shows positive adsorption [15], which suggests a rise in Q sorption and the deposition of new adsorbents. The heat constant \( b_T \) is positive according to the Temkin isotherm, indicating that the absorption is physical and exothermic [16], [17].

v. CONCLUSION

After this experimental study with the aim of testing our synthesized material prepared in the laboratory (Fh) with respect to water purification. We could conclude that at \( pH = 3.87 \), a temperature \( T= 25^\circ C \), a stirring speed \( v= 400 \) rpm, a grain diameter \( d= 0.125 \) mm, and a solid/liquid ratio \( r(S/L) = 10 \) mg/L, the maximum retention capacity is \( 40.9157 \) mg / g (The Langmuir isotherm) with a removal efficiency of 99.511% where equilibrium is reached after 15 minutes of contact.

ACKNOWLEDGMENT

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REFERENCES


